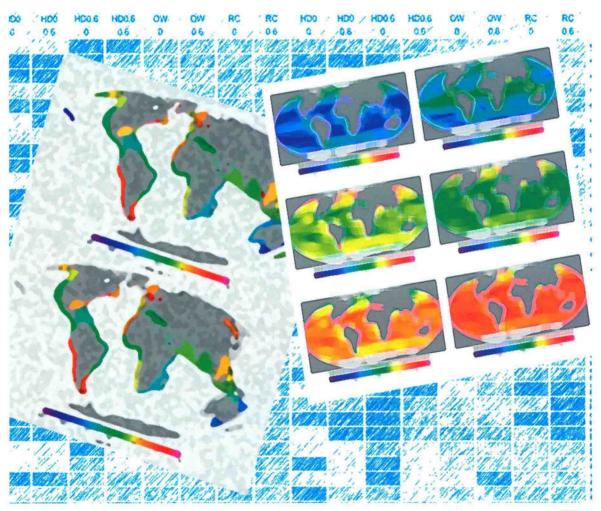
ISSN 2070-6065

DEVELOPING NEW APPROACHES TO GLOBAL STOCK STATUS ASSESSMENT AND FISHERY PRODUCTION POTENTIAL OF THE SEAS





The designations employed and the presentation of material in this information product do not imply the expression of any opinion whatsoever on the part of the Food and Agriculture Organization of the United Nations (FAO) concerning the legal or development status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. The mention of specific companies or products of manufacturers, whether or not these have been patented, does not imply that these have been endorsed or recommended by FAO in preference to others of a similar nature that are not mentioned.

The views expressed in this information product are those of the author(s) and do not necessarily reflect the views or policies of FAO.

ISBN 978-92-5-107992-8 (print) E-ISBN 978-92-5-107993-5 (PDF)

© FAO, 2014

FAO encourages the use, reproduction and dissemination of material in this information product. Except where otherwise indicated, material may be copied, downloaded and printed for private study, research and teaching purposes, or for use in non-commercial products or services, provided that appropriate acknowledgement of FAO as the source and copyright holder is given and that FAO's endorsement of users' views, products or services is not implied in any way.

All requests for translation and adaptation rights, and for resale and other commercial use rights should be made via www.fao.org/contact-us/licence-request or addressed to copyright@fao.org.

FAO information products are available on the FAO website (www.fao.org/publications) and can be purchased through publications-sales@fao.org.

DEVELOPING NEW APPROACHES TO GLOBAL STOCK STATUS ASSESSMENT AND FISHERY PRODUCTION POTENTIAL OF THE SEAS

Andrew A. Rosenberg

UCS, United States of America

Michael J. Fogarty

NFSC, NMFS, NOAA, United States of America

Andrew B. Cooper

SFU, Canada

Mark Dickey-Collas

ICES, Denmark

Elizabeth A. Fulton

CSIRO, Australia

Nicolás L. Gutiérrez

MSC, United Kingdom

Kimberly J.W. Hyde

NMFS, NOAA, United States of America

Kristin M. Kleisner

Sea Around Us Project, UBC, Canada

Trond Kristiansen

IMR, Norway

Catherine Longo

NCEAS, United States of America

Carolina V. Minte-Vera

UEMa Brazil, and IATTC, United States of America

Cóilín Minto

GMIT, Ireland

lago Mosqueira

EC JRC, IPSC, MAU, Italy

Giacomo Chato Osio

EC JRC, IPSC, MAU, Italy

Daniel Ovando

SFG, UC, United States of America

Elizabeth R. Selig

Gordon and Betty Moore Center for Science and Oceans, CI, United States of America

James T. Thorson

FRMD, NFSCm NMFS, NOAA, United States of America

Yimin Ye

FAO, Italy

FOOD AND AGRICULTURE ORGANIZATION OF THE UNITED NATIONS Rome, 2014

PREPARATION OF THIS DOCUMENT

FAO has been monitoring the state of the world's marine fish stocks since 1974, and it periodically produces the most authoritative report on the subject – *The State of World Fisheries and Aquaculture*. Information on the state of fishery sustainability is not only important for policy formulation, but also crucial to guide the fishing industry and its managers to develop effective harvest strategies. Moreover, sustainable fisheries require healthy ecosystems. To monitor ecosystem health, it is necessary to conduct ecosystem-level assessments that take into consideration both targeted and non-targeted species, interspecies interactions, and other factors that cannot be determined by looking at each stock in isolation. With these objectives, the Fisheries and Aquaculture Department of the FAO commissioned a study – Developing New Approaches to Global Stock Status Assessment and Fishery Production Potential of the Seas.

This circular presents the results of the study. It consists of two parts. Part 1 focuses on determining single-stock status and summarizes the results of simulation testing of four methods that can be applied to data-poor fisheries. Part 2 reports the results on the estimation of ecosystem-level production potentials based on satellite-based estimates of primary productivity.

Eighteen scientists around the world participated in this study: Andrew A. Rosenberg, Union of Concerned Scientists, United States of America; Michael J. Fogarty, Northeast Fisheries Science Center, National Marine Fisheries Services, National Oceanic and Atmospheric Administration, United States of America; Andrew B. Cooper, Simon Fraser University, Canada; Mark Dickey-Collas, International Council for the Exploration of the Sea, Denmark; Elizabeth A. Fulton, CSIRO, Australia; Nicolás L. Gutiérrez, Marine Stewardship Council, United Kingdom; Kimberly J.W. Hyde, National Marine Fisheries Services, National Oceanic and Atmospheric Administration, United States of America; Kristin M. Kleisner, Sea Around Us Project, University of British Columbia, Canada; Trond Kristiansen, Institute of Marine Research, Norway; Catherine Longo, National Center for Ecological Analysis and Synthesis, United States of America; Carolina V. Minte-Vera, Universidade Estadual de Maringá, Brazil, and Inter-American Tropical Tuna Commission, United States of America; Cóilín Minto, Galway-Mayo Institute of Technology, Ireland; Iago Mosqueira, European Commission Joint Research Centre, Institute for Protection and Security of the Citizen, Maritime Affairs Unit, Italy; Giacomo Chato Osio, European Commission Joint Research Centre, Institute for Protection and Security of the Citizen, Maritime Affairs Unit, Italy; Daniel Ovando, Sustainable Fisheries Group, University of California, Santa Barbara, United States of America; Elizabeth R. Selig, Betty and Gordon Moore Center for Science and Oceans, Conservation International, United States of America; James T. Thorson, Fisheries Resource and Monitoring Division, Northwest Fisheries Science Center, National Marine Fisheries Service, National Oceanic and Atmospheric Administration, United States of America; Yimin Ye, Food and Agriculture Organization of the United Nations, Italy.

Rosenberg, A.A., Fogarty, M.J., Cooper, A.B., Dickey-Collas, M., Fulton, E.A., Gutiérrez, N.L., Hyde, K.J.W., Kleisner, K.M., Kristiansen, T., Longo, C., Minte-Vera, C., Minto, C., Mosqueira, I., Chato Osio, G., Ovando, D., Selig, E.R., Thorson, J.T. & Ye, Y. 2014. Developing new approaches to global stock status assessment and fishery production potential of the seas. FAO Fisheries and Aquaculture Circular No. 1086. Rome, FAO. 175 pp.

ABSTRACT

Stock status is a key parameter for evaluating the sustainability of fishery resources and developing corresponding management plans. However, the majority of stocks are not assessed, often as a result of insufficient data and a lack of resources needed to execute formal stock assessments. The working group involved in this publication focused on two approaches to estimating fisheries status: one based on single-stock status, and the other based on ecosystem production.

For the single-stock status work, a fully factorial simulation testing framework was developed to assess four potential data-limited models. The results suggest that Catch-MSY, a catch-based method, was the best performer, although the different models performed similarly in many cases. Catch-MSY was more effective in estimating status over short time scales and could be particularly applicable for use in developing countries where data time series are often shorter. Harvest dynamics was the most important explanatory variable in determining performance, which emphasizes the importance of having accurate information on fishing effort and total removals.

For the ecosystem-level production analysis, the working group used satellite-based estimates of primary productivity by size classes and a more complete food web, which included more complete microbial pathways than earlier approaches. The working group also assembled estimates of ecological transfer efficiencies from a large number of energy flow network models to characterize uncertainty. The first-order estimates of fishery production potential indicated a potential yield of up to 180 million tonnes of fish, which could vary depending on the capacity to sustainably diversify the suite of species that are currently exploited. Planktivorous species provide the largest scope for growth. However, consideration of factors such as the ecological impact on other food web components, profitability of harvest operations, and marketability for these species must first be resolved. The realized production potential for planktivores may be much lower than their potential levels depending on the outcome of these considerations. The working group estimated that up to 50 million tonnes of benthic production could be potentially harvested, although this estimate is subject to similar constraints as those for planktivores. The greatest scope for growth in the benthic component may be found in the mariculture sector, subject to suitable environmental safeguards.

Ecosystem exploitation rates should not exceed 20–25 percent of available production, considering basic energetic constraints in marine ecosystems. Current harvest levels for benthivorous and piscivorous species (principally fish) exceeded these levels in higher-latitude ecosystems (subarctic-boreal and temperate) and were near or slightly below them in lower latitudes and upwelling systems. The estimates of the ratio of current catches to available production for planktivorous species are substantially lower, reflecting the production potential of currently underutilized species. However, targeted harvesting of selected planktivorous species does lead to relatively high exploitation rates for some species. Together, these results provide globally applicable methods for estimating fish stock status and fishery production potential.

ACKNOWLEDGEMENTS

The working group members would like to thank their respective organizations for supporting their work on the analyses and preparation of this report. These organizations include the Commonwealth Scientific and Industrial Research Organisation, EC Joint Research Center, Galway-Mayo Institute of Technology, Institute of Marine Research – Bergen, Norway, Inter-American Tropical Tuna Commission, International Council for the Exploration of the Sea, Marine Stewardship Council, National Center for Ecological Analysis and Synthesis, National Oceanic and Atmospheric Administration (NOAA) Northwest and Northeast Fisheries Science Centers and National Marine Fisheries Service, Oceanographic Institute at the University of São Paulo, Simon Fraser University, Universidade Estadual de Maringá, Sea Around Us Project at University of British Columbia, University of California – Santa Barbara, Union of Concerned Scientists, and Conservation International.

Special thanks go to Trevor Branch (University of Washington), Ray Hilborn (University of Washington), Hiroyuki Kurota (Seikai National Fisheries Research Institute, Fisheries Research Agency), Meredith Lopuch (Gordon and Betty Moore Foundation) and James Rising (Columbia University) for attending working group meetings and providing valuable feedback.

Gratitude also goes to the Norwegian Metacenter for Computational Science for its technical and logistical support during the working group's use of the supercomputer Hexagon for conducting the analyses.

Reg Watson (IMAS, UTAS) and Tilla Roy (LOCEAN-IPSL, Université Pierre et Marie Curie) are acknowledged for their help in calculating transfer efficiency statistics, as are Villy Christensen, Joe Buszkowski and Jeroen Steenbeek of the Fisheries Center, University of British Columbia, Canada, for their assistance in extracting Ecopath statistics.

Robert Gamble (NOAA Northeast Fisheries Science Center) assisted with the ecosystem production potential simulations, and Mariano Koen-Alonso (Northwest Atlantic Fisheries Centre, Fisheries and Oceans Canada) provided the WinBugs code adapted for those analyses.

The working group would also like to thank the University of Washington, the EC Joint Research Center, the New England Aquarium and the Marine Biological Laboratory for hosting working group meetings.

The working group was generously funded by FAO and by the Gordon and Betty Moore Foundation.

ABBREVIATIONS AND ACRONYMS

APE absolute proportional error

AR autoregression

BMSY biomass at maximum sustainable yield

CMSY catch-MSY COM catch only model

COM-SIR catch only model – sampling importance resampling

DB-SRA depletion-based stock reduction analysis

DCAC depletion-corrected average catch

EwE Ecopath with Ecosim HD harvest dynamics

ICES International Council for the Exploration of the Sea

ID initial depletion LH life history

LME large marine ecosystem

MAPE mean absolute proportional error

MPE mean proportional error

mPRM modified panel regression model

MSD maximum single density MSY maximum sustainable yield

MTL mean trophic level

NES LME Northeast US Continental Shelf LME OW one-way trip (a type of harvest dynamic)

PE proportional error

PPR primary production required PRM panel regression model

RC roller coaster (a type of harvest dynamic)

SIR sampling importance resampling SSCOM state–space catch only model

TL trophic level
TS time series length

VGPM Vertically Generalized Productivity Model

CONTENTS

	tion of this document	
	t	
Tables		
Figures		V11
	vledgements	
Abbrevi	iations and acronyms	X1
Overall	introduction	1
Part 1	Determining single-stock status	3
1.	Introduction	5
2.	Methods: evaluation of the performance of different models for estimating the status	of
2.	unassessed fisheries	
	Description of candidate models	7
	Modified panel regression model (mPRM)	7
	Catch-MSY model (CMSY)	9
	Catch only model – sampling importance resampling model (COM-SIR)	11
	State–space catch only model (SSCOM)	13
	Simulation framework and implementation	14
	Data generation framework using the Fisheries Library for R tools	14
	Deterministic independent variable design	16
	Stochastic independent variable design	
	Iterations	
	Estimation platform	21
	Data provided to method developers	21
	Evaluation of method performance	22
	Visualizations	22
	Performance metrics	22
	Regression trees	24
	Best performance visualization (tile plots)	24
3.	Results	27
	Overall perfomance	27
	Best performance	29
	Frequency of best performance	29
	Tile plots	29
	Performance maps	31
	Performance across models	
	Determinants of performance for each of the four assessment methods	39
	Modified panel regression model (mPRM)	39
	Catch-MSY (CMSY)	41
	Catch-only model (COM-SIR)	42
	State-space catch-only model (SSCOM)	44
4.	Discussion	47
Part 2	Fishery production potential	49
5.	Introduction	51
٥.	Methods and data sources	52
	Estimating primary production	
	Transfer efficiencies	57
	Benthic-mesozooplankton pathway	57
	Landings data	58
	Assignment of landings data to taxonomic groups	59
	Discard data	60

	Mean trophic level and species dominance of landings	60
	Catch-production ratios	
	Treating uncertainty	61
	Ecosystem-based exploitation reference levels	
6.	Results	
	Primary production	
	Transfer efficiencies	
	Production potential	67
	Landings and catch	
	Mean trophic level and dominance of landings	
	Yield efficiency index	
	Fishery production potential	
	Ratio of catch to available production	
7.	Discussion	
8.	References	
Appe	endix 1 Additional figures referred in Part I	91
	endix 2 Additional figures referred in Part II	

TABLES

1.	Overall model comparison	
2.	Summary of variables and model fits for the modified panel regression model (mPRM)	9
3.	A simple Schaefer production model and corresponding management parameters	
	relevant to the CMSY approach	10
4.	Resilience category by stock parameters (k parameter of the von Bertalanffy growth	
17.7	function, age at maturity T_{mat} , maximum age T_{max}) and corresponding prior distribution	
	for r	12
5.	Life-history specifications implemented in the simulations	
6.	Deterministic experimental design	
7.	Stochastic simulation experimental design factor levels	21
8.	Indices for results	23
8. 9.	Emergent variables estimated for biomass, catch and effort series	
		24
10.	Number of times a given method performed best as judged by mean absolute	20
	proportional error (MAPE) or mean proportional error (MPE) over all years	29
11.	Large marine ecosystems (LMEs) and designated ecotypes used in determining transfer	- 1
·	probability estimates	
12.	Designation of 29 functional groups from the Sea Around Us Project	60
13.	Ratio of microplankton to total production, nano-picoplankton coefficient of variation	100 424
	(CV) and microplankton CV by LME	
14.	Estimated potential yield by ecotype and functional group	77
15.	Estimated ratios of catch to available production for benthivorous and piscivorous fish	
	and for planktivorous fish	78
FIGUI	RES	
1.	Flow chart of stock simulation	16
2.	Deterministic simulation trajectories of rescaled stock biomass B/BMSY, fishing	10
2.	mortality (F) and resulting catch for each combination of the design given in Table 6	10
2		17
3.	Histograms of mean proportional error (MPE: left column) and mean absolute	20
	proportional error (MAPE: right column) across all stochastic simulations	28
4.	Best performance in each stochastic scenario according to the mean absolute	20
	proportional error (MAPE) statistic over the entire time series	30
5.	Relative performance in each stochastic scenario according to the mean absolute	Ξ.
	proportional error (MPE) statistic over the entire time series	31
6.	Performance maps by mean proportional error (MPE)	33
7.	Performance maps by mean absolute proportional error (MAPE)	
8.	Regression tree of proportional error (PE) across all methods for all years	36
9.	Regression tree of proportional error (PE) across all methods for the last five years	37
10.	Regression tree of absolute proportional error (APE) across all methods for all years	
11.	Regression tree of absolute proportional error (APE) across all methods for the last five	
	years	39
13.	Catch-MSY (CMSY) regression trees for: (A) proportional error (PE) last five years;	
15.	(B) PE all years of the catch time series; (C) absolute proportional error (APE) last five	
	years; (D) APE all years in the catch time series	42
14.	Catch-only model (COM-SIR) regression trees for: (A) proportional error (PE) last five	12
14.	years; (B) PE all years of the catch time series; (C) absolute proportional error (APE)	
		11
1.5	last five years; (D) APE all years in the catch time series	44
15.	State–space catch-only model (SSCOM) regression trees for: (A) proportional error	
	(PE) last five years; (B) PE all years of the catch time series; (C) absolute proportional	
	error (APE) last five years; (D) APE all years in the catch time series	
16.	Food web structure employed in this analysis	53
17.	Strata used in estimating global fishery production potential based on large marine	
	ecosystem boundaries	54

18.	Distribution patterns for total chlorophyll a, primary production, percentage	
	microplankton chlorophyll a, percentage microplankton primary production, percentage	-
19.	nano-picoplankton chlorophyll a, and percentage nano-picoplankton primary production	64
20.	Box plots of ecological transfer efficiencies	6/
20.	Estimated production levels in the absence of exploitation by functional group for LMEs represented in this study	68
21.	Trends in landings by ecotype for invertebrates, planktivores, and other fish	00
	(benthivores and piscivores) by ecotype	70
22.	Average landings by LME for the period 1998–2007(upper) and discard estimates	70
	(lower) from Kelleher (2005). Areas shaded in grey do not have available discard	
	estimates	71
23.	Trends in the mean trophic level of the catch (left panel) for all species and upper	/ 1
	trophic level species (right panel) by ecotype	73
24.	Trends in species dominance of the catch as measured by a modification of the Berger-	, _
	Parker diversity index	74
25.	Ratio of landings to total phytoplankton primary production (nano-picoplankton and	
	microplankton production) (upper), and the ratio of landings to microplankton primary	
	production (lower panel)	76
A1.1.	Deterministic simulation trajectories of rescaled stock biomass B/BMSY (black line),	
	and estimated B/BMSY for CMSY for each combination of the design given in Table 6.	
	Harvest dynamics levels (HD) and length of time series (TS) structure the columns,	
	initial depletion (ID) and life history (LH) structure the rows.	91
A1.2.	Deterministic simulation trajectories of rescaled stock biomass B/BMSY (black line),	
	and estimated B/BMSY for COM-SIR for each combination of the design given in	
	Table 6. Harvest dynamics levels (HD) and length of time series (TS) structure the	
	columns, initial depletion (ID) and life history (LH) structure the rows.	92
A1.3.	Deterministic simulation trajectories of rescaled stock biomass B/BMSY (black line),	
	and estimated B/BMSY for mPRM for each combination of the design given in	
	Table 6. Harvest dynamics levels (HD) and length of time series (TS) structure the	
	columns, initial depletion (ID) and life history (LH) structure the rows.	93
A1.4.	Deterministic simulation trajectories of rescaled stock biomass B/BMSY (black line),	
	and estimated B/BMSY for SSCOM for each combination of the design given in	
	Table 6. Harvest dynamics levels (HD) and length of time series (TS) structure the	
	columns, initial depletion (ID) and life history (LH) structure the rows.	94
A1.5.	Mean Proportional Error (MPE) estimated according to equation 21 for the 4 models	
	(CMSY, COM-SIR, mPRM and SSCOM) and each factor in the full factorial design.	
	Harvest dynamics levels (HD) and length of time series (TS) structure the columns,	
	initial depletion (ID) and life history (LH) structure the rows. The value of MPE	
	provides the average bias of the estimate (B/BMSY) and for example a value of 0.25	
	indicates that the model is overestimating B/BMSY by 25%	95
A1.6.	Iteration 1 of the stochastic simulation trajectories of rescaled stock biomass B/BMSY	
	(black line), and estimated B/BMSY for each model (CMSY, COM-SIR, mPRM and	
	SSCOM) for each combination of the design given in Table 8. Harvest dynamics levels	
	(HD) and length of time series (TS, only 60 years), level of recruitment variability	
	(sigmaR) and measurement error in catch (sigmaC) structure the columns. Initial	
	depletion (ID), life history (LH) and autocorrelation on recruitment residuals (AR)	
	structure the rows. The figure displays only the first iteration of each stochastic run out	
	of the 10 available for readability.	96

A1.7.	Iterations 1-10 of the stochastic simulation trajectories of rescaled stock biomass B/BMSY (black line) and estimated B/BMSY by each model (CMSY, COM-SIR,
	mPRM and SSCOM) for each combination of the design given in Table 8. Harvest
	dynamics levels (HD) and length of time series (TS, only 60 years), level of recruitment variability (sigmaR) and measurement error in catch (sigmaC) structure the columns.
	Initial depletion (ID), life history (LH) and autocorrelation on recruitment residuals
	(AR) structure the rows
A1.8.	Mean Proportional Error between true and estimated B/BMSY in all the 10 iterations of
0.000	the stochastic runs by recruitment variability (sigmaR), measurement error in catch
	(sigmaC), autocorrelation on recruitment residuals (AR) and model for all years
	available
A1.9.	Mean Proportional Error between true and estimated B/BMSY in all the 10 iterations of
	the stochastic runs by life history (LH), initial depletion (ID), harvest dynamics (HD),
	length of the time series (TS) and model for all years available
A1.10.	Regression tree of proportional error (PE) across all methods for all years for main
	factors and emergent properties variables. The top number in each box is the average PE
	for a set of simulation scenarios (i.e. the averaged PE across all methods and
	simulations was 0.29 or 29%). The numbers in the second row of the boxes list the
	number of data points and percentage of simulation scenarios in that set (i.e. the top box
	has 9350 scenarios representing 100% of the scenarios), and each box either has no
	boxes below it (i.e. it is a terminal node), or has two boxes below it (i.e. it has additional
	branching). The percentages in each box of a single tier sum to 100% (see Tables 7, 8
A 1 11	and 10 for factors, levels and emergent properties variables)
A1.11.	Regression tree of proportional error (PE) across all methods for the last five years for
A 1 12	main factors and emergent proprieties variables (see Figure A1.10 caption)
A1.12.	Regression tree of absolute proportional error (APE) across all methods for all years for
Δ1.13	main factors and emergent proprieties variables (see Figure A1.10 caption)
111.13.	Regression tree of absolute proportional error (APE) across all methods for the last five
A2.1.	years main factors and emergent proprieties variables (see Figure A1.10 caption)
112.11	1; e.g. miscellaneous fishes) to high (light blue, 6; e.g. species genus), in the landings by
j	decade in each Large Marine Ecosystem (LME)
A2.2 to	A2.70. The mean climatological (1998-2007) chlorophyll (CHL - left) and primary
	production (PP - right) on the top row; the mean microplankton and nano+picoplankton
	CHL and PP on the second row; and the percent CHL and PP attributed to the
	microplankton and nano+picoplankton size classes on the third for each LME and FAO
	region. The black line on each plot represents the LME boundary and the white line is
	the 300 m isobath. The composites also include climatological monthly and annual bar
	plots showing the seasonal and interannual variability of the size fractionated CHL and
	PP for each depth strata. Note, no depth strata data were calculated for the FAO
	subareas

OVERALL INTRODUCTION

Wild-capture fisheries provide a critical source of nutritional and economic benefits to people worldwide. In 2010, fisheries generated livelihoods and income for almost 38.5 million people (FAO, 2012) and currently fish provide approximately 3 billion people with almost 20 percent of their intake of animal protein. In the last half century, marine fisheries have been rapidly expanding and developing (Swartz *et al.*, 2010). Fishing fleets have also been increasing, both in number and extent, since the 1970s (Anticamara *et al.*, 2011; Watson *et al.*, 2013), although this growth has stabilized in the last decade (FAO, 2010). Concurrently, total landings increased from 16.8 million tonnes in 1950 to a peak of 86.4 million tonnes in 1996, but subsequently declined to 77.4 million tonnes in 2010 (FAO, 2012). With coastal populations projected to grow by 35 percent in the next 20 years, the demand for fisheries resources is likely to continue to increase. The combined intensification in both pressures on and demand for fisheries resources necessitates a broad understanding of the state of global fisheries to support policy formulation and the development of effective marine management.

In spite of their importance, it remains a major challenge to determine the status and potential production of wild-capture fish stocks. Managers and policy-makers need information on individual fish stocks to evaluate their status so that effective management strategies can be developed. At the same time, it is also necessary to undertake ecosystem-scale assessments that account for the interactions between stocks, the impact of fishing on non-target fish, and other factors that cannot be determined by looking at each stock in isolation.

Costello et al. (2012) estimated that more than 80 percent of the global catch comes from stocks that have not been formally assessed. Formal stock assessments require substantial data and resources to complete. Therefore, data-limited approaches are needed to assess the status of global fish stocks and to develop benchmarks for the fishery production potential of the oceans. The working group addressed these challenges using two approaches to estimate fisheries status: one based on singlestock status, and the other based on ecosystem production. The single-species work stream focused on evaluating the operational performance of different methods for estimating stock status within a simulation framework to evaluate their performance robustly. This simulation framework can also be used to examine the performance of other data-limited and data-rich approaches. The ecosystem production work stream was tasked with developing estimates of fishery production for each large marine ecosystem (LME) and FAO statistical area based on overall primary production in each area. This information allows for the extracted production to be compared with the estimated total production in an LME or FAO area, which is useful for developing food security policies, for effective marine stewardship, and for understanding the potential gains in fishery production from enhanced ocean management. Results from exploitation rates with estimated fishery production potential. enhanced ocean management. Results from both work streams can be used to compare current

There is always a trade-off beween risk and exploitation, and this study provides a suite of methods for evaluating fish stocks at greatest risk so that they can be prioritized for management and increased data collection. Estimating stock status and identifying regions that may be at risk for overexploitation are key components of moving to the sensuring sustainable exploitation. The work described in this report is an important step in investigating the performance of methods that can be used to estimate stock status. The results are not intended to provide direct advice to motivate management measures on specific fisheries, but to give an indication of the health of fish stocks and their production potential.

The approaches from the two work streams provide a more quantitative and consistent basis for evaluating global fish stock status than has previously been available. These estimates are vital for efforts to assess the health of marine ecosystems globally under data-limited situations.

www.earth.columbia.edu/news/2006/story07-11-06.php

PART 1 DETERMINING SINGLE-STOCK STATUS

w

1. INTRODUCTION

Managers and policy-makers need information on the status of individual fish stocks in order to manage marine fisheries resources sustainably, implement rebuilding plans for overfished species and increase production where possible. Formal stock assessments, often considered to be the gold standard in fisheries science, are available for a relatively small proportion of global stocks. Assessed stocks account for about 16 percent of harvested fish taxa (Ricard *et al.*, 2012), although the proportion of stocks assessed is likely to be lower for developing countries (Mora *et al.*, 2009). These assessments use all available data (e.g. catches, size and age distributions, surveys and tagging information) to quantify the rate of exploitation (F) in relation to that which is considered sustainable (F_{MSY}) and the relationship between historical and current stock biomass and the biomass that can produce maximum sustainable yield (MSY) (Branch *et al.*, 2011). This biomass ratio is commonly referred to as B/B_{MSY} .

In order to assess the status of fish stocks at the global level, FAO uses a combination of quantitative (formal) and qualitative stock assessments, using available information such as catch, abundance indices, spawning potential and age and size composition (FAO, 2012). In some cases, numerous types of data of varying quality are used for these assessments, but sometimes the only information that may be available is catch data (Branch $et\ al.$, 2011). These FAO assessments, which have been applied to 445 fish stocks since 1974, revealed that 30 percent of marine capture fisheries were overexploited and 57 percent of stocks were fully exploited in 2009 (FAO, 2012). Other research has estimated that 63 percent of assessed stocks require rebuilding to $B_{\rm MSY}$; therefore, greater efforts are needed to improve the health of fisheries (Worm $et\ al.$, 2009). In general, these global assessments provide an important overall picture of the health of fish stocks, but they are based only on a limited number of stocks. In some cases, these assessments do not provide target or limit reference points that can be used for management. However, both the formal assessment methods and the FAO assessments still omit many small stocks, many of which are vital for food security, especially in developing countries and small island nations.

The majority of commercially exploited species have never been assessed and no reference points have been established for them. Most methods for calculating stock status in data-limited fisheries rely solely on catch data. There has been considerable controversy over the use of catch data to estimate stock status for unassessed fisheries (Branch *et al.*, 2011; Pauly, Hilborn and Branch, 2013). Nonetheless, some studies show that small, unassessed stocks may be in poorer condition than suggested by global estimates of fisheries status, based largely on assessed stocks (Costello *et al.*, 2012; Froese *et al.*, 2012). Although formal stock assessments remain the standard for determining stock status and exploitation rates that can be used to inform management action, they will continue to be unfeasible for many of the world's fisheries because of the data and technical capacity required.

Determining stock status typically requires time-series information on historical removals (e.g. catch and discards), information on trends in abundance (e.g. catch per unit effort) and assumptions about the underlying processes that regulate or affect fish stocks (e.g. a production function such as a Schaefer production model, recruitment and/or assumptions about the economic drivers of fisheries). Only landings data exist for many data-limited stocks, which require additional assumptions, information and methods in order to estimate stock status.

There are both mechanistic and non-mechanistic methods that use only catch data to obtain a picture of stock status. Non-mechanistic approaches to assessing stock status include stock status plots, which use catch time series to assign development stages to individual stocks based on catch levels in relation to the maximum or peak catch of the time series (e.g. Froese and Kesner-Reyes, 2002; Pauly, 2007; Kleisner *et al.*, 2013). However, these methods have been criticized for their lack of mechanistic underpinnings (Branch *et al.*, 2011). In the United States of America, Congress tasked the National Marine Fisheries Service with the setting of annual catch limits and accountability measures for each managed fishery by fishing year 2010 for all stocks experiencing overfishing and by fishing year 2011 for all other stocks in the fishery (Berkson *et al.*, 2011). This mandate affected both datarich stocks for which traditional stock assessments could be conducted as well as data-limited stocks.