

Soil Science

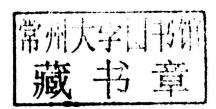
Environmental and Applied Aspects

Volume III

Brian Bechdal

Soil Science: Environmental and Applied Aspects Volume III

Edited by Brian Bechdal





Published by Callisto Reference, 106 Park Avenue, Suite 200, New York, NY 10016, USA www.callistoreference.com

Soil Science: Environmental and Applied Aspects Volume III Edited by Brian Bechdal

© 2015 Callisto Reference

International Standard Book Number: 978-1-63239-570-2 (Hardback)

This book contains information obtained from authentic and highly regarded sources. Copyright for all individual chapters remain with the respective authors as indicated. A wide variety of references are listed. Permission and sources are indicated; for detailed attributions, please refer to the permissions page. Reasonable efforts have been made to publish reliable data and information, but the authors, editors and publisher cannot assume any responsibility for the validity of all materials or the consequences of their use.

The publisher's policy is to use permanent paper from mills that operate a sustainable forestry policy. Furthermore, the publisher ensures that the text paper and cover boards used have met acceptable environmental accreditation standards.

Trademark Notice: Registered trademark of products or corporate names are used only for explanation and identification without intent to infringe.

Printed in China.

Soil Science: Environmental and Applied Aspects

Volume III

Preface

Environmental soil science is a wide area of study, which discusses the interaction of humans with the various layers of atmosphere such as pedosphere, biosphere, lithosphere and hydrosphere. Environmental soil science directly targets the fundamental as well as applied aspects such as surface water quality and buffers, vadose zone functions, land treatment and wastewater, stromwater, erosion related aspects, contamination of soil by metals and chemicals used in pesticides, remedies of soil contamination, degradation of soil, nutrients management techniques and developments and movement of different bacteria in the soil and water. Also included in this field are studies related to acid rains, land usage and global warming.

In contemporary times, the human civilisation is facing a lot of problems. Overpopulation happens to be a major cause of worry for a lot of countries. This is resulting in an extreme pressure on all our resources such as water, air, fuel etc. Feeding such a huge population is also an issue. In such a scenario, environmental soil science is the need of the hour. This is also because soil plays a crucial role in maintaining the various ecosystems which exist on Earth.

Instead of organizing the book into a pre-formatted table of contents with chapters, sections and then asking the authors to submit their respective chapters based on this frame, the authors were encouraged by the publisher to submit their chapters based on their area of expertise. The editor was then commissioned to examine the reading material and put it together as a book.

This effort would not have been possible without the kind cooperation of our contributors who patiently went through revisions of their chapters. I convey my heartfelt thanks to all the contributors and to the team at the publishing house for their encouragement and excellent technical assistance as and when required. I hope that this book will contribute significantly to both the basic and advanced group of readers, who are engaged in studies related to Environmental soil science.

Editor

Contents

	Preface	VII
Chapter 1	Effects of 24 Years of Conservation Tillage Systems on Soil Organic Carbon and Soil Productivity Kenneth R. Olson, Stephen A. Ebelhar and James M. Lang	1
Chapter 2	Effects of Paper-Mill Sludge as a Mulch versus Topsoil Incorporation on Potassium Uptake and the Grain Yield of Rain-Fed Wheat in a High Specific Surface Loess Soil with Illite Dominance in Clay Fraction S. Amini, S. A. R. Movahedi Naeini and K. Mashayekhi	11
Chapter 3	Suppression of <i>Bromus tectorum</i> L. by Established Perennial Grasses: Potential Mechanisms – Part One Robert R. Blank and Tye Morgan	21
Chapter 4	Acidification and Nitrogen Eutrophication of Austrian Forest Soils Robert Jandl, Stefan Smidt, Franz Mutsch, Alfred Fürst, Harald Zechmeister, Heidi Bauer and Thomas Dirnböck	30
Chapter 5	Soil Nutrient Availability, Plant Nutrient Uptake, and Wild Blueberry (Vaccinium angustifolium Ait.) Yield in Response to N-Viro Biosolids and Irrigation Applications Aitazaz A. Farooque, Farhat Abbas, Qamar U. Zaman, Ali Madani, David C. Percival and Muhammad Arshad	39
Chapter 6	Influence of Surface Biosolids Application on Infiltration Richard E. Zartman, Corey A. Moffet, David B. Wester, Ronald E. Sosebee, Ernest B. Fish and William F. Jaynes	46
Chapter 7	Nitrate-Nitrogen Leaching from Onion Bed under Furrow and Drip Irrigation Systems Parmodh Sharma, Manoj K. Shukla, Theodore W. Sammis and Pradip Adhikari	55
Chapter 8	A Critical Evaluation of Single Extractions from the SMT Program to Determine Trace Element Mobility in Sediments Valérie Cappuyns	72
Chapter 9	Soil Degradation-Induced Decline in Productivity of Sub-Saharan African Soils: The Prospects of Looking Downwards the Lowlands with the Sawah Ecotechnology Sunday E. Obalum, Mohammed M. Buri, John C. Nwite, Hermansah, Yoshinori Watanabe, Charles A. Igwe and Toshiyuki Wakatsuki	87

Chapter 10	Laboratory Experiments on the Effect of Microtopography on Soil-Water Movement: Spatial Variability in Wetting Front Movement Leif Sande and Xuefeng Chu	97
Chapter 11	Lifestyle Influence on the Content of Copper, Zinc and Rubidium in Wild Mushrooms J. A. Campos, J. A. De Toro, C. Pérez de los Reyes, J. A. Amorós and R. García-Moreno	105
Chapter 12	Soil Health Management under Hill Agroecosystem of North East India R. Saha, R. S. Chaudhary and J. Somasundaram	111
Chapter 13	Comparative Use of Soil Organic and Inorganic Amendments in Heavy Metals Stabilization Agustina Branzini and Marta Susana Zubillaga	120
Chapter 14	Fate in Soil of Flavonoids Released from White Clover (<i>Trifolium repens L.</i>) Sandra C. K. Carlsen, Hans A. Pedersen, Niels H. Spliid and Inge S. Fomsgaard	127
Chapter 15	Quantitative Analysis of Total Petroleum Hydrocarbons in Soils: Comparison between Reflectance Spectroscopy and Solvent Extraction by 3 Certified Laboratories Guy Schwartz, Eyal Ben-Dor and Gil Eshel	137
Chapter 16	Oilseed Meal Effects on the Emergence and Survival of Crop and Weed Species Katie L. Rothlisberger, Frank M. Hons, Terry J. Gentry and Scott A. Senseman	148
Chapter 17	Land Use Effects on Soil Quality Indicators: A Case Study of Abo-Wonsho Southern Ethiopia Awdenegest Moges, Melku Dagnachew and Fantaw Yimer	158
Chapter 18	Modelling Soil Water Retention for Weed Seed Germination Sensitivity to Water Potential W. John Bullied, Paul R. Bullock and Rene C. Van Acker	167
Chapter 19	Cadmium and Zinc Concentration in Grain of Durum Wheat in Relation to Phosphorus Fertilization, Crop Sequence and Tillage Management Xiaopeng Gao and Cynthia A. Grant	180
Chapter 20	Influence of Stabilized Biosolids Application on Availability of Phosphorus, Copper, and Zinc S. M. Shaheen, M. S. Shams, F. A. Elbehiry and S. M. Ibrahim	190
	Permissions	
	List of Contributors	

Effects of 24 Years of Conservation Tillage Systems on Soil Organic Carbon and Soil Productivity

Kenneth R. Olson, Stephen A. Ebelhar, and James M. Lang

¹ Department of Natural Resources and Environmental Sciences, University of Illinois, S-224 Turner Hall, 1102 S. Goodwin Avenue Urbana, IL 61801, USA

Correspondence should be addressed to Kenneth R. Olson; krolson@illinois.edu

Academic Editor: Philip J. White

The 24-year study was conducted in southern Illinois (USA) on land similar to that being removed from Conservation Reserve Program (CRP) to evaluate the effects of conservation tillage systems on: (1) amount and rates of soil organic carbon (SOC) storage and retention, (2) the long-term corn and soybean yields, and (3) maintenance and restoration of soil productivity of previously eroded soils. The no-till (NT) plots did store and retain 7.8 Mg C ha⁻¹ more and chisel plow (CP) -1.6 Mg C ha⁻¹ less SOC in the soil than moldboard plow (MP) during the 24 years. However, no SOC sequestration occurred in the sloping and eroding NT, CP, and MP plots since the SOC level of the plot area was greater at the start of the experiment than at the end. The NT plots actually lost a total of -1.2 Mg C ha⁻¹, the CP lost -9.9 Mg C ha⁻¹, and the MP lost -8.2 Mg C ha⁻¹ during the 24-year study. The long-term productivity of NT compared favorably with that of MP and CP systems.

1. Introduction

Conservation program was established to take highly erodible lands out of production. In the United States, the Food Security Act of 1985, the 1990, 1995, 2001, 2006, and 2011 Farm bills, and the Illinois T by 2000 Program have resulted in millions of hectares of erodible land previously in row crops being put into the CRP for 15 to 25 years. Any conversion of Conservation Reserve Program (CRP) land back to corn and soybean production could require the use of conservation tillage systems such as NT to meet soil erosion control standards. Evaluations of yield response of these conservation tillage systems over time are needed to assess returning this land to crop production, the effects on SOC storage and retention and crop yields.

Conservation tillage (defined as having 30% residue at the time of planting) can result in an increase in crop yield when compared with that of a moldboard plow system. Lawrence et al. [1] showed in a 4-year study in a semiarid environment in Australia that no-till had a higher crop yield than did

reduced till fallow or conventional till fallow. Wilhelm et al. [2] observed a positive linear response between yields of corn and soybean, and amount of residue applied to a notill system. Lueschen et al. [3], in a corn-soybean rotation in Minnesota, found an increase of 6.30 Mg ha⁻¹ in yield of the NT system above the MP system in a dry year. Kapusta et al. [4] studied the effects of tillage systems for 20 years and found equal corn yield for no-till, reduced till, and conventional tillage systems despite the lower plant population in no-till.

Maintaining crop residue on the soil surface [5, 6] can reduce the severity of erosion. At planting with chisel plowing, residue cover has to be 30 percent or more, but often much higher with no-till (usually 50% or more residue) due to minimum soil disturbance [7]. Lueschen et al. [3], for a corn-soybean rotation in Minnesota, observed 69 to 82, 49, and 10 percent of soybean residue cover on the soil surface after corn planting in no-tillage, chisel plow, and moldboard plow system plots, respectively.

The impact of tillage and cover crops on SOC sequestration (net increase) or loss has been the focus of many studies

² Dixon Springs Agricultural Center, Department of Crop Sciences, Simpson, IL 62959, USA

³ Department of NRES, University of Illinois, N-405 Turner Hall, 1102 S. Goodwin Avenue, Urbana, IL 61801, USA

since this management techniques are thought to contribute to atmospheric C loss or sequestration. SOC sequestration or retention has been shown to retain more SOC with decreasing disturbance or enhanced rotation diversity [8]. Many of the early studies found NT to have significantly higher SOC than MP and CP systems when the soils were only sampled to 15 or 30 cm depth [8, 9]. The NT plots have SOC concentrated in the top 30 cm, but was dispersed to greater depths in tilled plots [9, 10].

Decline of SOC content in agricultural systems and increased awareness of its importance to global C budgets has accelerated evaluations of land management impacts on soil C dynamics and storage [11]. Land use practices that may affect SOC sequestration include a switch to NT [12]. The impact of tillage on SOC sequestration (net increase) or loss has been the focus of many studies since these management techniques contribute to atmospheric C loss or sequestration. SOC sequestration or retention has been shown to occur with decreasing soil disturbance or enhanced crop rotation diversity.

Change in frequency and intensity of tillage practices alter the bulk density and soil organic matter in the soil profile. Mann [13] reported that the reduction in SOC content of soils having an initial content of between 20 and 50 g kg⁻¹ was 20 percent less after cultivation. He found that changes in SOC content were most pronounced during the first 20 years of cultivation. Also, changes in SOC storage were more variable in the upper 15 cm of soil than in the upper 30 cm. Varvel and Wilhelm [14] studied the use of conservation tillage systems for corn and soybean production and found soil organic carbon levels were maintained or even increased in all tillage system with the greatest increase obtained in systems with the least amount of soil disturbance which strongly support the adoption and use of conservation tillage systems for soil sustainability.

Franzluebbers and Follett [15] reported that SOC content of timberland and prairie soils declined with cultivation in North America. The rate of decline as a result of cultivation was 22% \pm 10% for the northeast; no value was reported for the central, 25% \pm 33% for southwest, and 36% \pm 29% for southeast. Ismail et al. [16] observed a decrease in SOC in the 0- to 30-cm silt loam layer of soil during the first 5 years, no change in the next 5 years, and an increase in SOC in the last 10 years in both NT and MP in comparison with sod plots. SOC was higher in NT than in MP. Hunt et al. [17] and Angers and Giroux [18] found NT systems increased SOC content, compared with MP and CP systems, in the top 5-cm layer of soils with a range of soil textures, including loamy sand, silt loam, and silty clay loam.

Mulvaney et al. [19] did not find an increase in soil organic matter build-up in response to increased crop residue input as a result of fertilizer N applications when a pretreatment SOC data were collected and used as the baseline to determine SOC change over time in long-term cereal cropping experiments. Khan et al. [20] found that the use of the comparison method with SOC content measurements only taken in the middle and at the end of a variable rate N fertilizer, crop yield, and SOC sequestration studies resulted in an overestimation of the magnitude and rate of SOC sequestration in response

to N fertilizer. Comparison studies with one treatment as the baseline or control should not be used to determine SOC sequestration, if soil samples are only collected and tested once during or at the end of study. Only experimental designs with pretreatment SOC measurements (baseline data) made before or at the start of long-term field studies should be used.

Olson [21] found that the SOC content of the MP treatment, used as baseline in comparison method studies, was not at steady state during the 20 yr tillage experiment at Dixon Springs, Illinois. In fact MP plots lost -15.2 Mg C ha⁻¹ from the root zone as a result of mixing, intensity of crop rotation, aeration, and eroded SOC rich sediments being transported off the plot area [22]. NT treatment only reduced the magnitude and rate of SOC loss over time. The pair comparison method used by many researchers with MP as baseline suggested +9.1 Mg C ha⁻¹ of SOC sequestration occurred during the 20 yr experiment [21]. However, the 1988 pretreatment baseline method did not validate the SOC sequestration value. At this site the sloping and eroding NT plot actually lost -6.8 Mg C ha⁻¹ during the 20-year study so no actually SOC sequestration occurred. The assumption that the MP was at a steady state, made by researchers [15, 23-27] using the comparison method with one year of SOC sampling near the end, was incorrect and resulted in an SOC sequestration finding that was invalid. Olson [21] findings suggest a pretreatment SOC baseline is essential in all tillage comparison studies to determine the amount and rate of SOC sequestration, steady state, or loss. A pretreatment SOC baseline was needed in these comparison studies when determining the amount and rate of SOC sequestration, storage, retention, or loss, especially on sloping and eroding soils with more intensive cropping rotations (more row crops and fewer years of forages) during the study than in previous

Shorter term tillage studies [28–30] were extended to 24 years with the objective of evaluating long-term tillage systems (NT, CP, and MP) effects on corn and soybean yields and the effects on the SOC storage or retention and the maintenance and restoration of soil productivity of previously eroded soils in southern Illinois. These SOC storage and retention values by tillage treatment at the end of the study will be compared to pretreatment SOC levels to determine the amount and rate of SOC sequestration or loss as a result of the 24 years of tillage treatments. The study was extended to show that NT system can be used instead of MP or CP systems to reduce soil erosion and maintain long-term crop yields.

2. Materials and Methods

A conservation tillage experiment was started in 1989 at the Dixon Springs Agricultural Research Center in southern Illinois. The soil at the study site was a moderately eroded phase of Grantsburg silt loam (fine-silty, mixed, mesic Typic Fragiudalf) [31] with an average depth of $64\pm 6\,\mathrm{cm}$ to a root-restricting fragipan. The area had an average slope gradient of 6 percent. Starting with corn in 1989, corn and soybean were grown in alternate years. The experimental design was two Complete Latin Squares and each square has three rows

and three columns [32] which allowed for randomization of the tillage treatments (NT, CP, and MP) both by row (block) and by column. This replication was used to control random variability in both directions. Each tillage treatment was randomized six times in 18 plots with a size of 9 m \times 12 m. The columns were initially separated with 6 m buffer strips of sod. Later the buffer strips were planted to NT corn and soybeans to reduce deer damage. An electric fence was later used to protect the crops in the plots. There was a 60 m wide filter strip between the plot area and the drainage way.

The implements used in each tillage system and depth of tillage were as follows: NT (John Deere No-Till planter with wavy coulters), CP (straight-shanked chisel plowed to 15 cm with disking to 5 cm), and MP (moldboard plowed to 15 cm with disking to 5 cm). In the spring of each year the MP and CP plots were moldboard and chisel plowed followed by 2 disking and planting. In odd years corn was planted at the seeding rate of 64,000 seeds ha⁻¹ with fertilizers of 218 kg ha⁻¹ N, 55 kg ha⁻¹ P, and 232 kg ha⁻¹ K. The tillage study did not focus on N fertilizer application in corn years since the rate was the same for both conventional (MP) and conservation (NT and CP) tillage systems.

The percentage surface residue was determined after planting by the line-transect method [33]. Plant population for the center 0.001 ha of each plot was determined by counting at 25 days after planting. The crop yield and plant population data from 1989 to 2012 were collected as part of this study. The soil loss rates were not measured directly. They were determined using RUSLE2 [34] and USLE [35] models that have been validated and widely used in USA by USDA, NRCS Soil Conservationists.

2.1. Field and Laboratory Methods. Soil samples were collected in September of 1988 (prior to the establishment of the tillage experiment in spring of 1989), in August of 2000 and in June of 2009, at depths of 0- to 5-, 5- to 15-, 15- to 30-, 30to 45-, 45- to 60-, and 60- to 75-cm for SOC determination. The sampling depth was limited due to the presence of a root restricting fragipan at a 64 ± 6 cm depth. Previous soil sampling found only trace amounts of SOC present below a 75 cm depth, probably from previous grass roots penetrating the fragipan along the prism faces. Four soil cores (3.2 cm diameter), one from near each of the four corners of the plot (1.5 m from adjacent, above or below plot, and 1.5 m from border strip), were obtained for each depth and composited by crumbling and mixing. The samples were air-dried and pulverized to pass through a 2 mm sieve prior to analysis. The SOC was determined after removal of un-decomposed plant residue using the modified acid-dichromate organic carbon procedure number 6A1 [36]. Field moist core bulk density was determined [36] using a Model 200 soil core sampler (5.6 cm in diameter and 6 cm high) manufactured by Soil Moisture Equipment Corp.

2.2. Statistical Methods. Statistical analysis for all parameters was performed using the procedures from Statistical Analysis System (SAS) computer software [37]. Analysis of variance and least square means of crop yield and SOC content were

performed by General Linear Model (GLM) procedures. Analysis of variance and least square means of crop yield and SOC content were performed by General Linear Model (GLM) procedures. An LSD procedure was used at the P=0.05 level to determine, if there were significant SOC differences between tillage treatments for the same date and depth.

3. Results

The NT system maintained a significantly greater amount of residue on the soil surface as compared with that of the CP, and MP systems at planting during each selected year (Table 1). Crop residue on the soil surface was greater with corn as previous crop, compared with that of soybean because of greater residue production from corn and slower rate of decomposition of corn residue [38] than soybean residue. On Grantsburg soil with 5–7 percent slopes, the estimated annual soil loss, determined with USLE and RUSLE2, was 8, 20, and 30 Mg ha⁻¹ with the NT, CP, and MP systems, respectively, (Table 1) [34, 35]. The greater the percentage of crop residue (Table 1) on the soil surface with the NT system protected the soil from erosion keeping it below the tolerance level of 8.4 Mg ha⁻¹ yr⁻¹ [35]. On the other hand, rill erosion was observed with the MP and CP systems as a result of fewer residues on soil surface compared with that of the NT system.

Rainfall data (30-year average growing season rainfall by month for the southeastern Illinois, USA) and 1989–2012 growing seasons are shown in Table 2. The 30-year average cumulative rainfall during April–September in southeastern Illinois was 64.4 cm. Seven years (1991, 1994, 1999, 2004, 2007, 2008, and 2012) could be characterized as dry years with a growing season rainfall of 43.3, 50.7, 47.7, 38.3, 44.4, 45.8, and 37.1 cm, respectively.

From 1989 to 2012, the MP system had greater plant populations than the other tillage systems in 6 of 24 years (Table 3). The NT system had greater plant population than the other tillage systems in 7 of 24 years while CP only had 1 of 24 years with greater plant populations than the other tillage systems. In 1989, 1996, 1998, 2002, and 2006, the NT had fewer plants per plot (Table 3) compared with the other tillage systems which was probably due to insufficient soilseed contact, lower germination, and greater soil strength in the NT system [28]. During 1990, 1996, 2002, 2006, and 2011 the high April and May rainfalls (Table 2) contributed to less plants per plot with the NT system compared with that of the MP system (Table 3). Better seed-soil contact with the MP system could have increased the germination compared with that of the NT system during 1996, 2002, and 2006 (Table 3). On the other hand, in 1994 and 2000 the plant population was greater with the NT treatment compared with that of the MP treatments, which could have been due to relatively greater water availability in the NT system compared with MP tillage system at planting. Twelve-year average plant population (Table 3) for corn and soybean was not statistically different from NT and MP systems.

In 2004, one of the driest years, the soybean yields were zero for all treatments (Table 4) since all plant available

Table 1: Effect of different tillage treatments on plant residue after planting and soil loss at Dixon Springs. Odd years have soybean residue and even years have corn residue.

Tillage	Residue present from previous crop (% cover)									Soil loss#
	1996	1997	1998	1999	2005	2006	2007	2011	2012	$(Mg ha^{-1})$
No-till	91a*	75a	95a	73a	85a	90a	78a	74a	88a	8c
Chisel plow	21b	18b	29b	21b	18b	28b	24b	20b	28b	20b
Moldboard plow	6c	6c	17c	5c	5c	10c	8c	6c	9c	30a

^{*}For each year means in a same column followed by the same letter is not significantly different at the P = 0.05 probability level.

TABLE 2: Rainfall data for 1989-2012 growing season at Dixon Springs in Southern Illinois.

Year	Rainfall (cm)								
1Ca1	April	May	June	July	August	September	Growing season		
1989	6.1	4.1	14.3	12.8	10.0	4.6	51.9		
1990	14.5	28.2	4.4	6.4	10.5	8.8	72.8		
1991	12.5	8.9	1.8	3.7	4.0	12.4	43.3		
1992	6.1	6.7	7.6	13.4	3.9	19.1	56.8		
1993	12.3	13.0	17.8	13.4	10.9	19.4	86.8		
1994	16.2	1.5	10.2	6.0	9.8	7.0	50.7		
1995	17.7	22.0	15.2	7.3	8.2	4.8	75,2		
1996	14.8	14.2	9.0	13.1	1.4	14.8	67.3		
1997	9.5	14.9	14.5	5.8	7.5	4.0	56.2		
1998	15.3	6.5	19.3	10.3	11.8	2.5	65.7		
1999	10.3	6.8	16.8	10.0	2.3	1.5	47.7		
2000	6.2	15.8	15.1	6.8	3.8	8.3	56.0		
2001	6.0	8.4	9.3	15.9	9.8	9.3	58.7		
2002	19.0	24.7	3.1	5.1	7.2	18.8	77.9		
2003	12.4	32.2	11.9	4.0	13.5	12.6	86.6		
2004	6.5	13.5	5.4	8.1	4.8	0.0	38.3		
2005	10.0	4.7	6.5	10.8	18.2	6.8	57.0		
2006	7.5	10.2	7.0	21.3	7.7	21.3	75.0		
2007	8.2	7.3	7.1	9.9	4.5	7.4	44.4		
2008	11.6	6.6	5.8	11.6	7.0	3.2	45.8		
2009	12.9	20.0	9.9	34.7	9.4	13.2	99.9		
2010	8.8	12.2	10.5	6.8	9.2	8.5	55.2		
2011	35.3	21.4	20.6	11.1	7.8	16.6	112.6		
2012	3.3	1.6	3.0	9.4	6.7	13.2	37.1		
1989-2012 average	11.0	12.1	9.8	9.8	7.9	9.5	59.8		
30-year average	12.3	13.9	10.3	10.2	8.3	9.4	64.4		

water above the fragipan was extracted from all treatments including the NT system. In 1994, another year of low rainfall, the soybean yields were low for all treatments, but NT yield (Table 4) was substantially greater than CP and MP yields. In 1999, the NT corn yield of NT system was significantly greater than CP and MP. The 24-year average rainfall for the April through September period was 59.8 cm which is equal to the 30-year average (Table 2). Years of 1990, 1993, 1995, 2002, 2003, 2006, 2009, and 2011 were considered wet years.

From 1989 to 2012, tillage affected crop yields in only 1989, 1994, 1999, 2001, 2002, 2006, 2009, 2011, and 2012 (Table 4). In 1994, 2002, and 2012, the NT system produced significantly greater soybean yield than with the CP and MP systems

partially due to greater plant population. Since 1994 and 2012 were dry years, and 2002 had a dry June to August period, the NT system could have provided more soil water to soybean at planting and later in the season compared with that of the other tillage systems. This enhanced soil water storage could have resulted in an improvement in nutrient availability and played an important role in greater soybean yields in 1994, 2002, and 2012 with the NT system as compared to MP and CP systems. Greater crop yield with the NT system than MP system in a dry year (not a drought) was also noted by Lueschen et al. [3]. Although the differences in soybean yield in 1996 were not significant by tillage treatment, the NT system had 8 and 16 percent greater yield than the MP and CP

^{*}Soil loss is calculated by Universal Soil Loss Equation (USLE) and Revised Universal Soil Loss Equation (RUSLE2).

Table 3: Effect of different tillage treatments on the corn and soybean populations during 1989-2012 at Dixon Springs.

	Plant population (1000 plants ha ⁻¹)								
Corn year	1989	1991	1993	1995	1997	1999			
NT	55.3b*	57.9a	51.4a	58.8ab	46.9a	69.2a			
CP	59.2ab	47.4b	54.1a	55.7b	51.9a	64.2ab			
MP	62.9a	52.2ab	52.6a	62.2a	51.9a	62.7b			
Year	2001	2003	2005	2007	2009	2011	12-year average		
NT	64.5a	59.5a	70.5b	68.8a	65.0a	75.7a	59.4a		
CP	56.5b	61.3a	76.0a	68.5a	65.5a	67.7ab	58.6a		
MP	65.9a	60.0a	69.5b	68.3a	64.0a	61.3b	60.2a		
Soybean year	1990	1992	1994	1996	1998	2000			
NT	191.0b*	344.0a	303.0a	263.0b	271.0b	422.0a			
CP	247.0a	335.0a	229.0b	277.0b	273.0b	398.0a			
MP	249.0a	343.0a	181.0c	309.0a	294.0a	405.0a			
Year	2002	2004	2006	2008	2010	2012	12-year average		
NT	293.0b	240.0a	320.0a	310.0b	294.0a	168.5a	284.4a		
CP	393.0a	228.0a	363.0b	390.0a	279.0a	125.6b	296.5a		
MP	420.0a	255.0a	388.0c	360.0ab	288.0a	176.0a	297.8a		

^{*}For each crop means in a same year followed by the same letter is not significantly different at the P = 0.05 probability level.

TABLE 4: Effect of different tillage treatments on corn and soybean yields during 1989-2012 at Dixon Springs.

Tillage				Crop yield (1	Mg ha ⁻¹)		
Corn year	1989	1991	1993	1995	1997	1999	2001
NT	8.99b*	6.57a	11.79a	11.60a	9.87a	8.12a	9.73b
CP	9.99b	6.10a	11.61a	11.55a	9.32a	6.78b	9.60b
MP	11.26a	6.60a	10.98a	10.37a	9.59a	6.98b	10.34a
Year	2003	2005	2007	2009	2011	12-yr ave.	% of MP yield
NT	6.67a	11.40a	6.46a	13.10a	6.13a	9.18a	96
CP	7.33a	11.82a	6.80a	13.60a	6.85b	9.24a	97
MP	7.82a	11.41a	7.33a	14.03b	7.68ab	9.53a	100
Soybean year	1990	1992	1994	1996	1998	2000	2002
NT	2.37a	3.74a	2.87a	2.63a	2.63a	2.32a	2.37a
CP	2.62a	3.46a	1.81b	2.27a	2.63a	2.38a	2.07a
MP	2.62a	3.65a	1.49b	2.43a	2.75a	2.32a	1.98b
Year	2004	2006	2008	2010	2012	12-yr ave.	% of NT yield
NT	0a	3.17ab	2.84a	2.94a	2.65a	2.54a	100
CP	0a	3.37a	2.84a	2.79a	1.96b	2.35a	93
MP	0a	3.10b	3.04a	2.88a	1.82b	2.34a	92

^{*}For each crop means in a same year followed by the same letter is not significantly different at the P = 0.05 probability level.

systems, respectively. In 2002, the NT soybean yield (Table 4) was significantly greater than MP when the plant population (Table 3) was significantly less. Year 2002 was considered a wet year; however, the combination of a dry period (between June and August of 2002) and the greater number of plants in MP resulted in less water available per plant. In 2004, there were no measureable soybean yields on any of the tillage treatments. Growing season rainfall totaled 38.3 cm which is 26 cm below the average. In addition, the fall 2003 and winter and spring of 2004 had below average rainfall and the 15 cm water storage capacity of the Grantsburg moderately eroded soils was not full on April 1 which caused the soybean yield to be zero on all treatments.

A tillage study was conducted to determine the amount and rate of SOC storage and retention in NT and CP tillage system when compared to MP system as the comparison baseline. The SOC levels were determined at the end of the 24-year study (Table 5). The MP SOC level (43.1 Mg C ha⁻¹ layer⁻¹) was significantly less for the 0–75 cm layer (root zone) than the NT (50.9 Mg C ha⁻¹ layer⁻¹) and similar to the CP SOC value (42.0 Mg C ha⁻¹ layer⁻¹). The NT plots did store and retain more SOC in the soil (7.8 Mg C ha⁻¹) and the CP plots less (–1.6 Mg C ha⁻¹) than MP. These SOC amounts were retained and stored in the soil and not decomposed and reemitted to the atmosphere as a

Comparison method NT or CP versus Paired comparison annual SOC retention Tillage treatment Depth cm June 2012 MP 24 yr total SOC retained above MP rate above MP the comparison baseline Mg C ha-1 layer-1 0 - 1526.3a** +7.5+0.31NT 15 - 7524.6a +0.01+0.30-75 (all) 50.9a +7.8+0.3220.3b** 0 - 15+1.5+0.06CP 15 - 7521.7a -2.6-0.110-75 (all) 42.0b -1.1-0.050 - 1518.8b MP 15 - 7524.3a

Table 5: Paired comparisons using MP as baseline to determine no-till (NT) and chisel plow (CP) effects after 24 years of tillage treatments on the volumetric SOC content change of the Grantsburg soil.

TABLE 6: Use of pretillage treatment 1988, 2000, and 2012 SOC during 24 years of tillage treatments.

43.1b

0-75 (all)

Tillage	Death	September 1988	August 2000	June 2012	
treatment	Depth cm	(pretreatment baseline)	Mg C ha ⁻¹ layer ⁻¹		
	0-15	28.5a**	26.8a**	26.3a**	
NT	15-75	23.6a	20.2a	24.6a	
	0-75 (all)	52.1a	47.0a	50.9a	
	0-15	28.4a	25.0a	20.3b	
CP	15-75	23.5a	18.7ab	21.7a	
	0-75 (all)	51.9a	43.7ab	42.0b	
	0-15	28.3a	19.9b	18.9b	
MP	15-75	23.1a	17.8b	24,3a	
	0-75 (all)	51.4a	37.7b	43.1b	

^{**}Means six replications with the same letter and in the same year and depth with a different tillage treatment are not significantly different at P = 0.05.

result of cultivation or in the transported sediment moved off of the plots.

Since the plot area was on 6% slopes and eroding, the NT plot SOC storage and retention amount and rate benefited from a smaller soil erosion rate of 8 Mg ha⁻¹ yr⁻¹ versus 20 Mg⁻¹ ha⁻¹ yr⁻¹ for CP and a soil erosion rate of 30 Mg⁻¹ ha⁻¹ yr⁻¹ for the MP plots (Table 1). The erosion rates were calculated using USLE and RUSLE2 [34, 35]. When MP loses SOC rich sediment from the plots and MP is the baseline, it has the effect of crediting NT (using comparison approach) with more SOC storage and retention.

Based on the paired comparison between NT and MP SOC in Table 5, it appears that NT stored and retained 7.8 (50.9–43.1) Mg C ha⁻¹ more SOC than the MP plots at the end of the 24-year study. Using the comparison method [15, 23–27] with SOC levels determined in the last year of the study and MP treatment as the baseline, the NT plots at Dixon Springs appear to have sequestered 7.8 Mg C ha⁻¹ of SOC in the soil. However, without knowing the SOC content in the

plot area prior to or at the start of the tillage experiment (pretreatment SOC baseline), it would not have been possible to determine if SOC sequestration actually occurred and at the annual SOC retention rate (0.32 (7.8/24) Mg C ha $^{-1}$ yr $^{-1}$) suggested in Table 5. These plots were sampled again and the SOC measured in August of 2000 and June of 2012 (Table 6) are lower than the initial September 1988 values for future NT, CP and MP plots sampled prior to tillage treatments being applied. The tillage plots (NT, CP and MP) lost SOC during the 24-year experiment and no SOC sequestration occurred on any of the tillage treatments. The NT plots actually lost a total of $-1.2\,\mathrm{Mg}\,\mathrm{C}\,\mathrm{ha}^{-1}$, the CP lost $-9.9\,\mathrm{Mg}\,\mathrm{C}\,\mathrm{ha}^{-1}$, and the MP lost $-8.2\,\mathrm{Mg}\,\mathrm{C}\,\mathrm{ha}^{-1}$ during the 24-year study (Table 7).

The annual loss in SOC in MgCha-1 yr-1 is shown in Table 8. The NT rate of SOC loss from the root zone during the first 12 years was $-0.42 \,\mathrm{Mg}\,\mathrm{C}\,\mathrm{ha}^{-1}\,\mathrm{yr}^{-1}$ but gained SOC at the rate of +0.33 Mg C ha⁻¹ yr⁻¹ during the last 12 years with an annual SOC loss rate for the 24 years of the tillage experiment was reduced to $-0.05 \,\mathrm{Mg} \,\mathrm{Cha}^{-1} \,\mathrm{yr}^{-1}$. The CP rate of SOC loss from the root zone during the first 12 years was $-0.68 \,\mathrm{Mg}\,\mathrm{C}\,\mathrm{ha}^{-1}\,\mathrm{yr}^{-1}$ and the loss rate decreased to -0.14 Mg C ha⁻¹ yr⁻¹ during the last 12 years with an annual SOC loss rate for the 24 years of the tillage experiment that was $-0.41 \,\mathrm{Mg} \,\mathrm{Cha}^{-1} \,\mathrm{yr}^{-1}$. The MP rate of SOC loss from the root zone (0-75 cm) during the first 12 years was -1.14 Mg Cha-1 yr-1 and gained +0.45 Mg Cha⁻¹ yr⁻¹ during the last 12 years making MP annual loss rate $-0.34 \,\mathrm{Mg} \,\mathrm{C} \,\mathrm{ha}^{-1} \,\mathrm{yr}^{-1}$ for the entire 24-year tillage experiment (Table 8). The higher rate of SOC loss from all tillage treatments during the first 12 years would have been as a result of the land use conversion and cultivation of the 15year-old sod covered plot area.

4. Discussion

The 12-year average corn yield and the 12-year soybean yields were not affected by tillage (Table 4). Twelve-year average MP soybean yield was 2 and 9% greater than NT and CP

^{**} Means six replications with the same letter and in the same year and depth with a different tillage treatment are not significantly different at P = 0.05.

TABLE 7: Use of pretillage treatment (1988) SOC values (sod) as baseline to determine the total SOC change in the first 12 years and during 24 years of tillage treatments. The 2000 SOC values for each plot were used as baseline to determine the total SOC content change during the last 12 years for the Grantsburg soil.

Tillage treatment (6 replications)	Depth cm	First 12-yr tillage effect on SOC total change by treatment using 1988 baseline	Last 12-yr tillage effect on SOC total change by treatment using 2000 baseline Mg C ha ⁻¹ layer ⁻¹ (% change)	24 year tillage effect on SOC total change by treatment using 1988 baseline
	0-15	-1.7 (-6)	-0.5 (-2)	-2.2 (-8)
NT	15-75	-3.4 (-14)	+4.4 (+22)	+1.0 (+4)
	0-75 (all)	-5.1 (-10)	+3.9 (+8)	-1.2 (-2)
	0-15	-3.4 (-12)	-4.7 (-19)	-8.1 (-28)
CP	15-75	-4.8(-20)	+3.0 (+16)	-1.8(-8)
	0-75 (all)	-8.2 (-16)	-1.7 (-4)	-9.9 (-19)
	0-15	-8.3 (-29)	-1.1 (-5)	-9.5 (-34)
MP	15-75	-5.3 (-23)	+6.5 (+36)	+1.2 (+5)
	0-75 (all)	-13.6 (-27)	+5.5 (+14)	-8.2 (-16)

Table 8: Use of pretreatment SOC baseline (1988) to determine the rate of SOC change During the first 12 years and for 24 years of tillage. A 2000 baseline was used for the last 12 years on the volumetric SOC content (Mg C ha⁻¹ layer⁻¹ yr⁻¹) of the Grantsburg soil.

Tillage treatment	Depth cm	1988 to 2000 tillage effect on SOC rate of change by treatment (baseline 1988)	2000 to 2012 tillage effect on SOC rate of change by treatment (baseline 2000) Mg C ha ⁻¹ layer ⁻¹ yr ⁻¹	24-year tillage effect on SOC rate of change by treatment (baseline 1988)
	0-15	-0.14	-0.04	-0.09
NT	15-75	-0.28	+0.36	+0.04
	0-75 (all)	-0.42	+0.33	-0.05
	0-15	-0.28	-0.39	-0.33
CP	15-75	-0.40	+0.25	-0.08
	0-75 (all)	-0.68	-0.14	-0.41
	0-15	-0.70	-0.09	-0.39
MP	15-75	-0.44	+0.54	+0.05
	0-75 (all)	-1.14	-0.45	-0.34

systems. The MP corn yields were 3% greater than for CP and NT systems as a result of significantly greater yields with MP system when planted into sod. At the beginning of the experiment, the MP system produced 21 and 11 percent higher yield compared with that of the NT and CP systems during 1989. The NT yields were lower in the early years of study but improved with the passage of time. The NT performance relative to MP and CP (Table 4) was better during dry years (1999) or years with extended dry period than wet years (Table 2), which was previously observed by Eckert [38]. The only year that CP treatment had the greatest yield was 2006 when compared to other tillage systems. The growing season rainfall that year was greater (75 cm) than average.

The NT crop yields were lower during the 3 early years (1989 to 1991) of the study but the NT system yielded as well as the MP system during the last 21 years of study. No-till yields were less than MP system in wet years (except 2002 with a dry period from June to August) but were greater in relatively dry years (Tables 2 and 4). The greater yields with the NT

system in relatively dry years (1994, 1999, 2002, and 2012) were probably due to the conservation of more soil water than the MP system. Chisel plow yields were less in wet years and greater in dry years as compared to MP system (Table 4).

In the 1860s, the plot area was under private ownership and was intensively cultivated until 1937 with some gully erosion evident in the 1937 air photographs. Land ownership changed from private to public at that time and plot area was then cultivated periodically until 1973 when it was in sod for 15 years. Clearly, the tillage effect on rate of SOC loss slowed or even started to gain over time (Table 6), but the NT, CP, and MP plots were still losing SOC (from pre-tillage treatment baseline) with no SOC being sequestered. Had the tillage experiment been established on a plot area which had been in row crops for the previous 15 years, it would appear that the rate of SOC loss to atmosphere and in transported sediment from MP, CP, and NT would have been reduced (Table 6). Tillage (NT and MP) treatments were starting to gain SOC in the last 12 years of the 24-year tillage study. In the NT, the

SOC additions could be as a result of the annual additions of plant roots in topsoil and subsoil and crop residue being added on the soil surface. In the case of MP, the SOC annual additions included plant roots into the topsoil and subsoil and the incorporation of both roots and plant tops into the topsoil layer.

The SOC content of the MP (comparison baseline) did not appear to be at or reach a steady state during the 24-year experiment since the SOC losses from water erosion, moldboard plow tillage, and mixing, some disturbance during planting and subsequent nitrogen injection in corn years and aeration was greater than any SOC sequestration or net SOC storage gain or mineralization gain in MP plots. Consequently, the SOC storage and retention reported in Table 5 for the NT system could not be verified as SOC sequestration using the pretreatment SOC values measured before during the 24-year tillage experiment and no SOC sequestration occurred on the NT plots. The same is true for the CP and MP plots.

If one were to compare the pretreatment SOC values for the NT, CP and MP (Table 8) with those measured at the end of the 24-years of a tillage experiment, one would find that SOC sequestration did not occur and that the NT lost $-0.05 \,\mathrm{Mg} \,\mathrm{C} \,\mathrm{ha}^{-1} \,\mathrm{yr}^{-1}$, the CP $-0.41 \,\mathrm{Mg} \,\mathrm{C} \,\mathrm{ha}^{-1} \,\mathrm{yr}^{-1}$, and the MP lost $-0.34 \,\mathrm{Mg} \,\mathrm{Cha}^{-1} \,\mathrm{yr}^{-1}$. The NT SOC sequestration rate of -0.05 Mg C ha⁻¹ yr⁻¹ (actually a SOC loss) is inconsistent with the central USA soil carbon sequestration rate of $0.40 \pm 0.61 \,\mathrm{Mg} \,\mathrm{Cha}^{-1} \,\mathrm{year}^{-1}$ [23] and the southeast rate of $0.45 \pm 0.04 \,\mathrm{Mg} \,\mathrm{Cha}^{-1} \,\mathrm{year}^{-1}$ [26] and inconsistent with a published national rate of 0.50 Mg (or MT) Cha⁻¹ year⁻¹ [27]. No SOC sequestration occurred on the 24 yr tillage study since the SOC levels of NT, CP, and MP were higher before the tillage treatment was applied than at the end of the study. The primary reason for SOC loss was the transport of carbon rich sediment from the plots by water erosion (responsible for 54 to 120% of the SOC storage loss in 24 years), moldboard plow or chisel plowing and mixing, some disturbance even on NT plots during planting and subsequent nitrogen injection in corn years, aeration, and mineralization.

Soil carbon sequestration (SOC) is the process of transferring carbon dioxide from the atmosphere into the soil through plants, plant residues, and other organic solids which are stored or retained as part of the soil organic matter (humus) [22]. The retention time of sequestered carbon in the soil can range from short-term (not immediately released back to atmosphere) to long-term (millennia) storage. The sequestered carbon process should increase the net soil organic carbon storage during and at the end of a study to above pretreatment baseline levels and result in a net reduction in the carbon dioxide levels in atmosphere. Olson [21] amended the definition of SOC sequestration to include meaningful boundaries to be used to measure actual changes in a specific part of a terrestrial (soil) pool. The proposed definition of soil organic carbon sequestration is the "process of transferring CO₂ from the atmosphere into the soil of a land unit through unit plants, plant residues, and other organic solids, which are stored or retained in the unit as part of the soil organic matter (humus)".

Much of the literature [15, 23-27] suggested that paired comparisons between conservation tillage and conventional tillage at the end of a short- or long-term study can be used to determine SOC sequestration rate. Researchers assumed that the conservation and conventional tillage plots have the same SOC level at the start of the tillage experiment, the conventional tillage plots maintained the SOC over time (at steady state), and any increase in SOC of the conservation tillage treatment above the conventional tillage plot at end of study represented the amount of SOC sequestered by the conservation tillage system. However, these studies often lacked or did not report a pretreatment baseline SOC content in the plot areas collected prior to or at the start of the tillage experiment. Without such pretreatment baseline data, the SOC sequestration magnitude and rate findings cannot be verified [21].

If SOC sequestered in conservation tillage plots at the end of the experiment is higher than the initial SOC of the plot area then SOC sequestration would have occurred in the conservation tillage (NT) plots and at the measured rate. Alternatively, if the conservation tillage plots did not have more SOC content at the end of the study than at the start of the experiment (pretreatment baseline), then no SOC content sequestration occurred and the SOC sequestration rate is not correct. This long-term study was conducted on a plot area that was previously in sod for 15 years and on a 6 percent slope. The Grantsburg soils were previously eroded as a result of cropland use between 1860 and 1973. If the comparison approach is used, the projected SOC storage and retention (not sequestration) rate for NT would have been 0.32 Mg C ha⁻¹ yr⁻¹ (Table 5) which was less than the published regional SOC sequestration rate averages (0.40 ± 0.61 Mg Cha⁻¹ yr⁻¹) for the north central USA region [23]. Regional studies included both nearly level (flat) and sloping and eroding plot areas.

Since the SOC content in the Dixon Springs plot area was measured before the establishment of the tillage experiment (pretreatment baseline), it was possible to determine that the NT plots had lost -1.2 Mg C ha⁻¹ (or 2%) and the MP and CP plots had loss SOC -8.3 Mg Cha-1 (or 16%) and -9.9 Mg Cha⁻¹ (or 19%), respectively over the 24 years. Contributing substantially to the above SOC losses from the tillage plots as a result of SOC rich sediment being transport off of the plots for the 24 yr study was 2.4 Mg C ha⁻¹ for NT, 5.3 Mg Cha⁻¹ for CP, and 7.2 Mg Cha⁻¹ for MP. The NT plots did retain more SOC than MP plots; however, they did not sequester SOC. NT treatment only reduced the magnitude and rate of SOC loss over time. It is true that if a farmer had decided to use MP instead of NT the amount of SOC retained in the soil after 24-years of NT treatment would be 7.8 Mg C ha⁻¹ greater than after 24 years of MP treatment and therefore the greenhouse gas emissions from the SOC in the NT plots would be less than from the SOC in the MP plots (but still greater than if the plot area had remained idle or in sod).